

1 ***Post-disturbance haulout behaviour of harbour seals***

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31 **1. ABSTRACT**

- 32 1. We investigated the impact of anthropogenic activity associated with marine
33 renewable developments on harbour seals (*Phoca vitulina*) using controlled
34 disturbance trials.
- 35 2. Hauled out seals were approached by boat until all seals had entered the water
36 and this was repeated approximately every three days (weather permitting). The
37 time taken for seal counts to return to pre-disturbance levels was determined by
38 monitoring haulout sites using time-lapse photography.
- 39 3. Mean post-disturbance counts of hauled out seals returned to 52% (95%CI 35-
40 69%) of pre-disturbance counts within 30 minutes. However, mean counts only
41 returned to 94% (95%CI 55-132%) of pre-disturbance counts after four hours.
- 42 4. Eight seals were tagged with GPS phone tags to provide information on haulout
43 location and at-sea movements, allowing investigation of how disturbance may
44 influence haulout site choice and seal distribution.
- 45 5. Telemetry tagged seals displayed a high degree of haulout site fidelity.
46 Disturbance trials did not have a significant effect on the probability of seals
47 moving to a different haulout site.
- 48 6. When seals hauled out again within the same low tide period after disturbance
49 trials, the proportion of time spent hauled out was high indicating that when
50 seals are motivated to haulout they will do so despite past disturbance.
51 Motivation to haul out more on disturbance trial days was not linked to a cyclic
52 pattern of hauling out more over consecutive low tide periods.
- 53 7. As there was no large scale re-distribution after disturbance we suggest that
54 monitoring effort to determine the effects of short-term increases in levels of
55 disturbance caused by boat activity can be spatially localized. However, where
56 disturbance is likely to be longer-term or impact on important haulout sites for
57 breeding and/or moulting, monitoring may be required over a larger
58 geographical area.

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62 **KEYWORDS**

63 Anthropogenic activity, coastal, intertidal, behaviour, disturbance, mammals,
64 hydropower, phocid, renewable energy

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69 **2. INTRODUCTION**

70 The spatial and temporal overlap of marine habitats used by humans and marine
71 mammals is an issue of growing concern. Development of marine renewable energy
72 technology has led to increased levels of construction activity in the marine environment
73 that, in some cases, results in avoidance behaviour by marine mammals (Dahne et al.,
74 2013; Russell et al., 2016). This could lead to barrier effects that exclude animals from
75 areas regularly used for foraging and, in the case of seals, for hauling out. The
76 commitment of many countries to an increased reliance on marine renewable energy is
77 likely to lead to an increase in the development of technologies that potentially have a
78 negative impact on the marine environment. Of those technologies, tidal turbine arrays
79 are expected to become an established technique with several projects already at an
80 advanced stage (Lewis et al., 2011). Tidal turbine deployments are best suited to areas
81 where tidal streams are restricted topographically resulting in faster currents and
82 therefore a higher energy yield (Lawn, 2009), meaning that sites identified for
83 deployment are often close to shore. For species where marine habitat use overlaps
84 with inshore areas identified as suitable for tidal turbine deployments there is a need to
85 assess the impact on these species before the construction phase commences.

86

87 In the UK a number of tidal turbine projects are under development (Uihlein & Magagna,
88 2016). Permitting such developments requires a realistic assessment of their likely
89 impact on marine mammals. Research aimed at meeting these requirements has
90 quantified the effects of marine renewables solely within the marine environment itself
91 (Hastie et al., 2015; Hastie et al. 2017; Thompson, Onoufriou, Brownlow & Morris, 2016;
92 Wilson, Benjamins & Elliott, 2013). However, the habitat use of harbour seals (*Phoca
93 vitulina*) includes terrestrial haulout sites that are important at various stages of their
94 annual life cycle (Thompson, Fedak, McConnell & Nicholas, 1989). Harbour seals have
95 been shown to forage relatively close inshore in some areas (Sharples, Moss, Patterson
96 & Hammond, 2012; Thompson et al., 1996) and display a high degree of site fidelity for
97 particular haulout sites (Cordes & Thompson, 2015; Dietz, Teilmann, Andersen, Riget &
98 Olsen, 2013). Inshore developments are likely to spatially and temporally overlap with
99 habitat regularly used by harbour seals. There is therefore potential for the construction,
100 operational and decommissioning phases of inshore marine renewable developments
101 to affect how harbour seals use the area in the vicinity of those developments for transit,
102 foraging and hauling out.

103

104 Several studies have described the normal haulout pattern of harbour seals in relation
105 to environmental conditions (Grellier, Thompson & Corpe, 1996; Watts, 1992), tidal
106 state (Pauli & Terhune, 1987), diurnal activity (Russell et al., 2015; Watts, 1996) and
107 seasonal events such as the breeding and moult periods (Thompson et al., 1989). Where
108 a novel stimulus resulting from increased anthropogenic activity creates a behavioural
109 response that results in a deviation from that normal haulout pattern, animals can be
110 considered to have been disturbed. Previous studies looking at the causes of disturbance
111 of seals at haulout sites have focused on the causes of disturbance, looking into factors
112 such as the distance at which seals are disturbed by boats (Jansen, Boveng, Dahle &

113 Bengston, 2010), the type of boat activity that causes disturbance (Johnson & Acevedo-
114 Gutierrez, 2007) and disturbance by pedestrians (Osinga, Nussbaum, Brakefield, & Haes,
115 2012). However, having identified the causes of disturbance it is important to then
116 quantify the consequences in terms of behavioural changes. UK harbour seals are listed
117 as a protected species under Annex II of the European Habitats Directive. Particularly in
118 Scotland, Section 117 of the Marine (Scotland) Act 2010 states that it is an offence to
119 “intentionally or recklessly harass seals” at designated haulout sites. Understanding
120 what happens when a normal haulout pattern is disrupted by anthropogenic activity is
121 key to meeting monitoring requirements aimed at mitigating against the impact of
122 disturbance on seals.

123

124 Changes in levels of anthropogenic activity have been shown previously to alter the
125 haulout behaviour of harbour seals. For example, Henry & Hammill (2001) suggest that
126 increased leisure activity increased the number of occasions harbour seals flushed into
127 the water in Métis Bay, Canada. Similarly, Lonergan, Duck, Moss, Morris & Thompson
128 (2013) suggest that harbour seals on the west coast of Scotland haul out less at the
129 weekends as opposed to during weekdays. Harbour seals may also switch to a nocturnal
130 haulout pattern to avoid hauling out during the day when daytime anthropogenic
131 activity is high (London, Hoef, Jeffries, Lance & Boveng, 2012). Increased anthropogenic
132 activity can therefore be a factor when observing broad-scale changes in the timing and
133 frequency with which harbour seals haul out. As well as quantifying how seal activity is
134 affected at particular sites it is also important to determine whether or not seals transit
135 from one location to another in response to disturbance (Andersen, Teilmann, Dietz,
136 Schmidt & Miller, 2014) which may require monitoring over a larger spatial scale. This is
137 particularly true where disturbance results in animals being displaced from sites
138 designated for protection. The spatial scale of monitoring should necessarily include the
139 area in the immediate vicinity of any proposed marine renewable development but also
140 the geographical range over which it is determined that increased anthropogenic activity
141 may have an effect.

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143 One such development is the tidal turbine array granted permission for deployment in
144 the Sound of Islay, Scotland (Paterson, Russell, Wu, McConnell & Thompson, 2015;
145 Sparling, 2013). In terms of impact on marine mammals, this site is of particular
146 importance due to its proximity to the South East Islay Skerries SAC designated to
147 protect harbour seals that use the site to haul out throughout the year. Harbour seals in
148 this area are known to transit between the South East Islay Skerries SAC and the Sound
149 of Islay in which the tidal turbine array is to be deployed. As well as being a regular
150 transit route for seals there are a number of harbour seal haulout sites within the Sound
151 of Islay that are in close proximity to the proposed development (Paterson et al., 2015;
152 Sparling, 2013). Here we describe a study to assess the behavioural responses of harbour
153 seals to disturbance from boat traffic within the Sound of Islay. By implementing a series
154 of controlled disturbance trials where hauled out seals were repeatedly approached by
155 boat until they entered the water, this study quantifies the associated effects in terms
156 of changes in haulout patterns and haulout site fidelity. The results are used to

157 determine the spatial extent of monitoring required when assessing changes in harbour
158 seal haulout behaviour affected by boat disturbance.

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163 **3. METHODS**

164 **Study sites**

165 Two sites on the eastern shore of Islay (55°45'N, 06°16'W), an island off the west coast
166 of Scotland, were chosen as focal haulout sites for this study (Figure 1). Both haulout
167 locations, Rubha Bhoraraic (RBR) and Bunnahabhain (BHN), were determined to be
168 regularly used by harbour seals based on aerial survey data collected between 1990 and
169 2009 and a previous telemetry-based study of seal movements and haulout site use in
170 2011 and 2012 (Sparling, 2013). Those data also indicated that RBR and BHN are two of
171 the most frequently used harbour seal haulout sites close to a proposed tidal turbine
172 development within the Sound of Islay. None of the haulout sites targeted in disturbance
173 trials were on the list of sites designated to provide additional protection from
174 intentional or reckless harassment of seals under Section 117 of the Marine (Scotland)
175 Act 2010. RBR and BHN are tidally influenced haulout sites with tidal ranges of between
176 1.0m and 1.5m during neap tides and 0.3m and 2.2m during spring tides. This results in
177 both haulout sites being fully submerged during spring high tides and remaining partially
178 available during neap high tides.

179

180 **Monitoring focal haulout sites using remote cameras**

181 Time-lapse photographs were collected at one minute intervals at both BHN and RBR.
182 Both camera systems consisted of two Canon EOS 1100 DSLR cameras in a single
183 weatherproof housing. Each housing had one camera equipped with an 18-55mm lens
184 and the other with a 70-300mm lens. This system provided both a wider scale view of
185 vessel activity around the haulout site to record when disturbance events occurred and
186 a narrower view more focused on the haulout site itself to determine the number of
187 seals hauled out. When conditions permitted, counts were made each minute between
188 the hours of 04:00 and 22:00 each day. Counts of seals were grouped by month and
189 each seal count was assigned values for three tidal state variables based on the time
190 since low water (LW), tidal height at the time of counting and tidal amplitude (difference
191 between predicted high water (HW) and LW heights). Counts were designated as high
192 tide or low tide if they occurred more or less than three hours from LW respectively and
193 as spring tide or neap tide if the tidal amplitude was in the upper or lower half of the
194 amplitude range for that spring/neap cycle. Tidal values were taken from the nearest
195 local reference port (Port Askaig; 3.8km from both RBR and BHN sites) in the POLTIPS
196 tidal prediction package (version 3.2.4, Proudman Oceanographic Laboratory).

197

198 ***Disturbance of seals at focal haulout sites***

199 Harbour seals at the South East Islay Skerries SAC and other haulout sites around Islay
200 generally come ashore on small rocky outcrops that are only accessible by boat. The type
201 of disturbance most relevant to the proposed tidal turbine array at the Sound of Islay is
202 a higher than normal exposure to boat traffic during the construction, operational and
203 decommissioning phases. To simulate this type of increased anthropogenic activity,
204 experimental disturbance trials were carried out by approaching hauled out seals in a
205 4.3m RIB at a speed of five knots. Direct approaches were initiated at a distance of
206 approximately 300m and continued in a straight line until the haulout site was reached
207 and all seals were flushed into the water. Seals were approached at an angle that
208 provided the clearest line of sight between animals on the haulout and the approaching
209 boat. Disturbance of seals from their haulout site was restricted to one trial per day,
210 approximately two hours before low tide to allow time for animals to haul out again
211 within the same low tide period. Over the study period disturbance trials were carried
212 out on a three-day cycle, dependent on navigable conditions. Disturbance trials at focal
213 haulout sites were carried out whenever harbour seals were present, regardless of
214 whether any of the telemetry tagged seals were present. The number of seals hauled
215 out at the point of disturbance was used as a reference for estimating the percentage
216 recovery of hauled out seals after disturbance trials.

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218 **GPS/GSM phone tag deployment**

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220 In April 2014 eight adult female harbour seals were captured for telemetry tag
221 deployment at either RBR (n = 2) or BHN (n = 6). Seals were captured using a pop-up net
222 that could be deployed underwater at low tide and remotely triggered to float to the
223 surface when seals hauled out in front of it during a subsequent low tide. Seals were
224 weighed before being anaesthetized with a 1:1 combination of Tiletamine and
225 Zolazepam (Zoletil® 100). GPS/GSM phone tags (McConnell, Fedak, Hooker & Patterson,
226 2010) were then glued to the seals' fur using Loctite® 422 Instant Adhesive. All
227 procedures were carried out under Home Office Animals (Scientific Procedures) Act
228 licence number 60/4009.

229 GPS/GSM phone tags were programmed to record an animal as having hauled out when
230 the on-board wet/dry sensor was continuously dry for >10 minutes. GPS location fixes
231 were collected while seals were at sea as well as on land. Data collected by the tag were
232 sent back to SMRU via the GSM mobile phone network providing daily updates of the
233 most recent location fixes. Recent movement patterns were used to assess the
234 likelihood of a seal being at or close to haulout sites in the study area. Table 1 gives the
235 latitude and longitude of all haulout locations used by telemetry tagged seals during this
236 study. Figures 1 and 2 present those locations on maps to show the relative distance
237 between visited haulout sites.

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239

240 ***Disturbance of telemetry tagged seals***

241 Telemetry tagged seals were disturbed into the water at RBR and BHN when present on
242 trial days. However, in order to maximize the number of disturbance trials with
243 telemetry tagged seals the recent movements of seals were examined to identify
244 additional sites where telemetry tagged seals were likely to be hauled out. Those sites
245 were then visited approximately two hours before low tide and wherever telemetry
246 tagged seals were found the same method of approach by boat used at RBR and BHN
247 was applied.

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250 ***Haulout transition rates***

251 Haulout events recorded by the tag were assigned a location. When multiple GPS points
252 were recorded while a seal was hauled out the median coordinates were used to assign
253 the location of the haulout event. However, the time series of GPS fixes were irregular
254 and so there were haul out events during which no locations were obtained. When this
255 happened, an approximate location was calculated using linear interpolation of GPS
256 locations immediately preceding and immediately following the haulout event. In
257 parallel, a list was accumulated of 'known haulout' sites that had been visited at some
258 time by these or previously tagged seals. Note that haulouts (as defined by >10 minutes
259 continuous dry rule) occasionally occurred at sea due to animals resting at the surface
260 for prolonged periods with the tag exposed to the air. Such at-sea (here defined as >2km
261 from the shore) haulouts were omitted from this analysis. In this study a haulout event
262 was defined as having ended when the tags were wet for >10 minutes. An animal was
263 then defined as being on a trip. The location and time until a subsequent haulout event
264 then determined if an animal had returned to the same haulout site or transited to a
265 different haulout site and in what timeframe either of these events occurred.

266 The first week's data were excluded from the final dataset. This allowed time for any
267 behavioural changes associated with seals being captured to return to normal
268 (McKnight, 2011). All statistical analyses were carried out using the statistics package R
269 (R Development Core Team, 2014). The modelling approach used examined how the
270 probability of hauling out at a different haulout site was influenced by time of year, site
271 fidelity, whether or not seals hauled out on the same or a subsequent low tide between
272 trips, and whether or not a disturbance event had taken place. The response variable
273 transition was binary in that having embarked on a trip to sea seals either transited from
274 one haulout site to another (1) or returned to the same haulout site (0). Both Julian day
275 and site fidelity were included as smooth terms (thin plate regression splines) to capture
276 the non-linear effects of both variables. Julian day was included to test for seasonal
277 effects. Levels of site fidelity vary by individual through time thus the percentage of
278 haulout events in the previous week that were at the current haulout location was used

279 as a measure of site fidelity for that particular site. Whether or not seals hauled out
280 during the same or a subsequent low tide period was included as a factor to determine
281 to what extent seals enter the water then haul out again at the same site or switch
282 haulout sites within a single low tide. In the context of disturbance this is relevant in that
283 once disturbed into the water, seals could either; (i) haul out within the same low tide
284 period at the same haulout site, (ii) haul out again within the same low tide period at a
285 different haulout site, (iii) haul out on a subsequent low tide period at the same haulout
286 site, or (iv) haul out on a subsequent low tide period at a different haulout site.
287 Disturbance was included as a factor, defined as whether or not seals were flushed into
288 the water during a haulout event while carrying out controlled disturbance trials. The
289 full model also included an interaction between site fidelity and tidal cycle because the
290 effect of site fidelity on transition probability may depend on whether animals haul out
291 in the same or a subsequent low tide period. A Generalized Additive Mixed Model
292 (GAMM) framework within the mgcv library (Wood, 2004) was used for analyses. An
293 AR1 correlation structure from the nlme library (Pinheiro, Bates, DebRoy, Sarkar & R
294 Core Team, 2018) was incorporated to account for temporal autocorrelation within
295 individuals. The error family used in all models was binomial. Backward model selection
296 was carried out using Akaike's Information Criterion (AIC) selection.

297

298 ***Proportion of time hauled out over consecutive low tide periods***

299 To investigate whether seals were in a cyclic pattern of hauling out more or less when
300 disturbance trials were carried out, the proportion of time spent hauled out was
301 compared over the consecutive low tide periods preceding, during and following
302 disturbance. To do this a generalized linear mixed effects model approach was
303 implemented using the R package glmmTMB (Brooks et al., 2017). The full model
304 included the fixed factors consecutive low tide period (three levels; pre-disturbance,
305 disturbance, post-disturbance), seal reaction i.e. whether they hauled out again within
306 the same or during a subsequent low tide after disturbance trials (two levels; same,
307 different) and the interaction between the two. To account for non-independence of
308 data within individuals, individual ID was included as a random effect. Binomial model
309 selection was performed by backwards selection using AIC. *Post hoc* pairwise
310 comparisons to investigate differences in the proportion of time spent hauled out over
311 consecutive low tide periods were made using the R package lsmeans (Lenth, 2016).

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318 **4. RESULTS**

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320 **Monitoring focal haulout sites using remote cameras**

321 Time-lapse photographs were collected between 23/04/2014 and 22/07/2014. Mean
322 counts relative to low tide are summarized for BHN and RBR in Figures 3 and 4
323 respectively. For both sites combined, the overall mean number of seals hauled out was
324 significantly lower (t-test, $p < 0.01$) at spring high tide ($\bar{x} = 0.12$, SE = 0.05) compared
325 with at neap high tide ($\bar{x} = 1.00$, SE = 0.16). This is due to the largest spring high tides
326 resulting in haulout sites occasionally being completely submerged resulting in
327 increased counts of zero. Mean seal counts were not significantly different ($p = 0.33$) at
328 spring low tide ($\bar{x} = 1.45$, SE = 0.24) compared with at neap low tide ($\bar{x} = 1.79$, SE =
329 0.20). During neap high tides haulout sites still remained available to seals to haul out
330 but were much reduced in size compared to during low tides. Mean seal counts at neap
331 high tide were significantly lower than at neap low tides ($p < 0.01$) and lower, but not
332 significantly ($p = 0.12$), than at spring low tide.

333

334 ***Disturbance of seals at focal haulout sites***

335 The first controlled disturbance trials were carried out on 26/05/2014 and continued on
336 a three-day cycle thereafter, dependent on navigable weather conditions, until
337 15/07/2014.

338 At BHN a total of 17 disturbance trials were recorded using time-lapse photography with
339 an average of 3.3 days (range = 3 to 4, SE = 0.11) between trials. Figure 5 shows the
340 mean number of seals counted post-disturbance expressed as a percentage of the
341 original number of seals counted immediately before disturbance trials were carried
342 out. Mean pre-disturbance counts of seals at BHN were 3.25 (range = 1 to 5, SE = 0.53)
343 during spring tides and 3.89 (range = 2 to 7, SE = 0.48) during neap tides. The difference
344 in means of pre-disturbance counts of seals during spring and neap tides at BHN were
345 not different (t-test, $p = 0.39$) and so data were pooled when assessing recovery rate.
346 Other than the telemetry tagged seals, it was not possible to identify individual seals to
347 determine whether seals that hauled out post-disturbance were the same as those
348 present before disturbance. It may therefore be that post-disturbance counts were
349 inflated by the presence of non-disturbed seals. However, the number of seals on the
350 haulout returned to 52% (95%CI 35-69%) of pre-disturbance levels within 30 minutes
351 and 94% (95%CI 55-132%) of pre-disturbance numbers within four hours. Beyond that
352 time, the influence of the rising tide caused mean counts to decline. Time-lapse
353 photography showed that BHN was regularly used as a haulout site throughout this
354 study with zero seal counts on only two days in May, three in June and one in July. Seals
355 were therefore available for disturbance trials on almost every occasion the site was
356 visited.

357 At RBR a total of 10 disturbance trials were recorded with an average of 6.2 (range = 3
358 to 27 days, SE = 2.62) days between trials. The low number of trials recorded at RBR
359 compared with BHN was due to the fact that on several occasions when disturbance
360 trials were due to be carried out there were no animals on the haulout. On each occasion
361 when disturbance trials were undertaken only one seal was present at RBR. There were
362 11 days in May, 17 days in June and 11 days in July when time-lapse photography
363 showed there to be no seals hauled out at RBR at low tide. This low level of haulout
364 activity was also reflected in the telemetry data as only one of the telemetry tagged
365 animals in this study visited RBR after April. In all 10 disturbance trials at RBR no seals
366 hauled out again within 30 minutes post-disturbance.

367

368 **GPS/GSM phone tag deployment**

369 GPS/GSM phone tag deployment resulted in a total of 626 days of data collected from
370 eight adult female harbour seals. The mean duration of tag deployment was 78 days
371 (range = 41 to 107, SE = 6.98). For all animals there was a total of 634 haulout events
372 separated by more than 10 minutes with a mean trip duration of 18.54 hours (range =
373 0.17 to 267.17, SE = 1.15) between haulouts. Overall, 16 haulout sites were used
374 throughout the study with individual seals using a mean of five haulout sites (range = 3
375 to 9, SE = 0.77). The mean duration of haulout events not including those in which
376 disturbance trials were conducted was 5.2 hours (SE = 0.28) (Table 1).

377

378 ***Disturbance of telemetry tagged seals***

379 A total of 15 disturbance trials were carried out at sites with telemetry tagged seals
380 between 29/05/2014 and 16/07/2014, by which time the majority of GPS/GSM phone
381 tags had ceased transmitting data. On four occasions more than one telemetry tagged
382 seal was present at the site where disturbance trials took place resulting in 22 seal
383 disturbance events overall. Table 2 summarizes the haulout sites at which telemetry
384 tagged seals were disturbed and whether they hauled out within the same or on a
385 subsequent low tide period. In 13 of the trials, animals hauled out again within the same
386 low tide period. On 12 of those occasions seals returned to the same haulout location
387 and only once did a seal transit to a different haulout site within the same low tide
388 period. The remaining nine seal disturbance events resulted in seals starting a trip that
389 included at least one high tide period. On eight of these occasions, seals later returned
390 to the haulout site from which they departed and on only one occasion did a seal haul
391 out at a different site on a subsequent low tide.

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395 ***Haulout transition rates***

396 A total of 626 trips (at sea periods of over 10 minutes) were identified. Trips that resulted
397 in seals transiting from one haulout site to another totalled 162 (26%) and had a mean
398 trip duration of 34.10 hours (SE = 4.58). The remaining 464 trips that resulted in seals
399 returning to the same site had a mean trip duration of 14.25 hours (SE = 0.95). Overall,
400 the maximum trip duration undertaken by a seal was 11 days. However, 75% of trip
401 durations lasted less than 24 hours. For trips that resulted in a transition to another
402 haulout site, the mean number of times that seals had hauled out at that site in the
403 previous week was 2.6 (SE = 0.27) compared to 7.2 (SE = 0.28) when it was a return trip.

404 For the 162 trips that resulted in a transition, only 13 were transitions to a different site
405 within the same low tide period. Two of those trips occurred after a controlled
406 disturbance trial. The remaining 149 trips were transitions that occurred on a
407 subsequent low tide which suggested that seals travelling from one haulout site to
408 another were more likely to do so having been at sea for a longer period. Of the 464
409 return trips, 51 occurred within the same low tide period. Additionally, 11 of these trips
410 were undertaken directly after controlled disturbance trials. The remaining 413 return
411 trips occurred on a subsequent low tide period. Overall, whether trips were transitions
412 or returns, 90% were separated by at least one high tide period.

413 Backwards AIC selection on the initial full model (AIC = 3010) resulted in Julian day and
414 disturbance being excluded as explanatory variables. This suggests that the probability
415 of seals transiting from one haulout site to another did not significantly change over the
416 course of this study and that overall, transition probability was not significantly affected
417 by disturbance trials. The final model (AIC = 2808) retained the interaction between site
418 fidelity and tidal cycle with significant smooths fitted separately for transition
419 probability dependent on level of site fidelity for seals hauling out during the same ($p =$
420 0.02) or a subsequent ($p < 0.01$) low tide period. An AR1 correlation structure that
421 accounted for temporal autocorrelation within individuals was also retained in the final
422 model. Figure 6 shows that probability of transition decreased as seals' fidelity for the
423 site at which they were hauled out increased. However, when a trip in between two
424 haulout events included at least one high tide period, the probability of transition was
425 generally higher than if that trip was completed within the same low tide period.

426

427 ***Proportion of time hauled out over consecutive low tide periods***

428 In the full model (AIC = -65) the inclusion of the interaction of consecutive low tide
429 period and seal reaction did not improve the model fit and was therefore excluded.
430 When treating consecutive low tide period and seal reaction as separate explanatory
431 factors seal reaction was also found not to improve the model fit and was therefore not
432 retained. This resulted in only consecutive low tide period being included as an
433 explanatory variable in the final model (AIC = -66). Post-hoc pairwise comparisons of the
434 proportion of time spent hauled out over consecutive low tide periods showed that
435 during low tide periods when seals were disturbed they spent a higher proportion of

436 time hauled out compared to the low tide periods immediately preceding ($p < 0.01$) and
437 following ($p = 0.04$) disturbance trials. The proportion of time spent hauled out during
438 the low tide periods prior to and following disturbance trials were not different from
439 one another ($p = 0.64$). GLMM model predictions of the mean proportion of time spent
440 hauled out during low tide periods with 95% confidence intervals are summarised in
441 Figure 7.

442

443 **5. DISCUSSION**

444 The normal haulout pattern of seals in this study was similar to previous studies in which
445 a high tidal range resulted in preferred haulout sites only being periodically available (Da
446 Silva & Terhune, 1988; Granquist & Hauksson, 2016; Pauli & Terhune, 1987). Time-lapse
447 photography revealed that focal haulout sites in the vicinity of the proposed tidal
448 turbine array in the Sound of Islay were either completely submerged or greatly reduced
449 in size during spring high tide and neap high tide respectively. During spring tides,
450 disturbance events that cause seals to enter the water from their haulout site reduces
451 the amount of time available to haul out at that site within a low tide period. Post-
452 disturbance there is a finite time within which disturbed seals can haul out again before
453 the flooding tide makes the haulout site unavailable. Also, the high site fidelity shown
454 by seals during this study meant that seals were unlikely to move to alternative locations
455 that continued to be available at high tide. During neap high tides, focal haulout sites
456 were not fully submerged and remained available to seals in a much smaller capacity.
457 This resulted in smaller groups occasionally hauling out over the high tide period.
458 However, seals in the Sound of Islay hauled out in larger numbers over low tides when
459 the time and space available for hauling out was maximal compared with high tide when
460 space on haulout sites was limited or non-existent. This effect was more pronounced
461 during spring tides compared with neap tides.

462

463 The purpose of this study was to quantify behavioural changes associated with a
464 stimulus that would have been perceived as novel by animals, such as that created
465 during a marine renewable development. The type and frequency of disturbance seals
466 were exposed to during trials represents the extreme scenario that all approaches by
467 boat result in seals flushing from the haulout site. However, it is important to note that
468 approaches by boats associated with a tidal turbine deployment in the Sound of Islay
469 are unlikely be in such close proximity to the haulout site and so are not expected to
470 elicit the same response seen during disturbance trials. Indeed, time-lapse photography
471 indicated that at the two focal haulout sites no boat activity other than that used during
472 trials caused animals to flush into the water suggesting that seals in the Sound of Islay
473 are not currently exposed to disturbance by boats that would be of concern. It may be
474 that harbour seals in the Sound of Islay are already habituated to existing levels of boat
475 traffic as observed in other studies (Johnson & Acevedo-Gutierrez, 2007; Mathews et
476 al., 2016).

477

478 In the present study, individuals on focal haulout sites could not be identified using time-
479 lapse photography meaning that it was not possible to quantify whether the response
480 of individual seals changed over time as a result of habituation. However, disturbance
481 trials that included telemetry tagged seals showed that no behavioural change was
482 observed over time in terms of the use of preferred haulout sites. This was despite there
483 being alternative haulout sites around Islay that seals could travel to. Site faithfulness of
484 seals remained high throughout even in the presence of a novel stimulus that
485 periodically caused those individuals to flush from their haulouts.

486
487 Disturbance trials were implemented at focal haulout sites two hours before low tide to
488 allow time within that same low tide period for the numbers of seals to recover towards
489 the original hauled out group size. It may have been the case that seals hauling out post-
490 disturbance were different to those exposed to disturbance trials. However, given the
491 high levels of site fidelity shown by seals during this study it is likely that at least some
492 of the seals returning to the haulout site post-disturbance were the same as those pre-
493 disturbance. At the more regularly used site (BHN) the rate of recovery was relatively
494 quick as haulout numbers returned to half that of pre-disturbance levels in the first half
495 hour post-disturbance. Haulout numbers did not approach the original state until
496 approximately four hours later indicating that time spent hauled out over the low tide
497 period would have been reduced for some individuals. The mean haulout duration of
498 undisturbed telemetry tagged seals was 5.2 hours (SE = 0.19) which is in line with a
499 previous study at the same site (Cunningham *et al.*, 2009). Seals flushed into the water
500 during disturbance trials would not have had this time available to them for a continuous
501 haulout either between the end of the preceding high tide or the start of the following
502 high tide and the point at which disturbance trials took place. Suryan & Harvey (1998)
503 showed that groups of hauled out harbour seals exposed to disturbance events that
504 caused them to enter the water were more likely to return to their original number when
505 disturbance events occurred earlier, compared to later in the low tide period.
506 Disturbance trials in the present study may therefore have had a greater impact in terms
507 of whether seals returned to haul out or not had they been implemented at a later stage
508 of the low tide period.

509 The timing of the implementation of disturbance trials may generally have affected the
510 results of this study dependent on how motivated seals were to haul out at particular
511 times. Despite haulouts being interrupted, the proportion of time spent hauled out was
512 higher over low tide periods when disturbance trials were implemented compared to
513 during the immediately preceding and following low tide periods. When the reaction of
514 seals to disturbance trials was to haul out again within the same low tide period
515 motivation to haul out could already have been higher on those occasions. However, it
516 does not seem that this was linked to seals spending a higher proportion of time hauled
517 out over consecutive low tide periods. Seal reaction was not retained as an explanatory
518 variable during model selection suggesting that when seals hauled out again after
519 disturbance trials and therefore spent a higher proportion of time hauled out, that
520 decision was not motivated by a cyclic pattern of hauling out more over consecutive low
521 tides. Motivation to haul out can also be associated with changes in at-sea activities in

522 the lead up to a haulout (Thompson et al., 1989). Trip duration at sea prior to the haulout
523 period in which disturbance trials were implemented was highly variable, making it
524 difficult to associate motivation to haul out with the need to rest after longer periods at
525 sea or indeed with any cyclic pattern of at-sea activity. The variability in trip duration
526 leading up to a haulout period was evident both when the response of animals to
527 disturbance trials was to haul out again within the same low tide ($\bar{x} = 19.38$, SE = 6.49,
528 range = 1.49 to 68.48) or on a subsequent low tide ($\bar{x} = 30.39$, SE = 11.40, range = 1.16
529 to 110.38). Regardless, when seals hauled out again after being disturbed they were
530 motivated on those occasions to do so, with the net effect of disturbance being to
531 disrupt what may otherwise have been a continuous haulout.

532 Reducing the time available for seals to haul out or increasing the frequency with which
533 animals enter the water has important implications for periods when harbour seals haul
534 out more often, such as during the breeding season (Cordes & Thompson, 2015) or
535 during the moult (Thompson et al., 1989). Being disturbed into the water may be
536 particularly important for pups that risk hypothermia due to lower insulation compared
537 with adults. Harbour seal pups primarily suckle while on land (Renouf & Diemand, 1984)
538 and where haulout sites are only tidally available there is a limited amount of time
539 during which suckling events can occur (Reijnders, 1981). If the frequency with which
540 mother pup pairs are forced into the water is sufficiently high then this could have
541 energetic consequences for pups (Jansen et al., 2010). A negative energy balance will
542 affect mass at weaning which has been shown to correlate with reduced over-winter
543 survival in young harbour seals (Harding, Fujiwara, Axberg & Harkonen, 2005). There
544 may also be consequences for adult seals that are moulting as repeated immersion due
545 to disturbance will increase heat loss and reduce skin temperature which may impede
546 the growth of new hair (Paterson *et al.* 2012). Disturbance trials in this study were not
547 undertaken at sites identified as being important habitat for breeding or moulting and
548 so a tidal turbine deployment in the Sound of Islay is not likely to have a significant
549 impact on harbour seals during these periods. However, it is essential that assessments
550 of the impact of marine renewable deployments on haulout behaviour of harbour seals
551 take into account proximity to habitat used by seals at different times of the year.

552
553 Disturbance trials of the type and frequency carried out during this study did not
554 influence the transit of seals from one haulout site to another. This resulted in
555 disturbance not being an explanatory factor in the final transition model. Site fidelity
556 was retained showing that seals were more likely to make a transition from a haulout if
557 they had visited it infrequently in the previous week. This agrees with other harbour seal
558 studies in which fidelity for particular haulout sites was high (Cordes & Thompson, 2015;
559 Dietz et al., 2013). Seals embarking on trips that included at least one high tide period
560 were also more likely to switch haulout sites. This suggests that unavailability of
561 preferred haulout sites during high tides and/or longer trip duration influenced
562 transition probability. Where seals showed a high level of fidelity for a particular site in
563 the previous week the probability of transition was very low regardless of the tidal cycle
564 when seals hauled out again. Andersen et al. (2014) also found that harbour seals in the
565 Kattegat Sea showed a high degree of site fidelity when exposed to repeated

566 disturbance trials. However, small tidal amplitudes meant that haulout sites were
567 available to seals at all states of the tide post-disturbance, meaning the option of
568 returning to the original haulout site was always possible. Large tidal amplitudes at the
569 Sound of Islay caused preferred haulout sites to become unavailable, presenting a
570 temporal and spatial challenge to seals disturbed from haulout sites. Despite preferred
571 haulout sites having limited availability in each tidal cycle and even with repeated
572 exposure to disturbance, seals still chose to return to preferred haulout sites when they
573 were available.

574 Our results show that at least on the time-scale of a few months harbour seals do not
575 make large scale movements between haulout sites in response to boat disturbance.
576 The level of disturbance in this study was likely greater than from the proposed tidal
577 development or from other anthropogenic sources in the Sound of Islay at the present
578 time. We therefore expect that increased anthropogenic activity associated with marine
579 renewables in the Sound of Islay would not change the distribution of harbour seals in
580 the short-term. However, previous studies have shown that harbour seals can be
581 displaced from haulout sites when exposure to anthropogenic activity is continued over
582 several years (Becker, Press & Allen, 2009; Becker, Press & Allen, 2011). Monitoring
583 harbour seal haulout sites during and beyond the construction phase of a marine
584 renewable development may therefore be necessary. In the case of harbour seals in the
585 Sound of Islay, the nearest habitat identified as being important for breeding and
586 moulting is the South East Islay Skerries SAC. In all SACs designated as such by the
587 presence of harbour seals, general advice to the public to avoid disturbing seals includes
588 not approaching animals to the point that they flush from their haulouts and maintaining
589 an appropriate distance when using recreational boats (Scottish Marine Wildlife
590 Watching Code, 2017). Dependent on the expected level of disturbance and how
591 habituated animals are to boat traffic this general advice may also be sufficient for
592 marine renewable developments. None of the telemetry tagged seals in this study
593 visited the South East Islay Skerries SAC and for these animals at least the effect of
594 disturbance was spatially localized to the haulout sites outside the SAC. Nevertheless,
595 where disturbance events associated with future marine renewable developments
596 exceed the type, frequency or duration imposed during this study, monitoring harbour
597 seal haulout behaviour may be required on a larger geographical and temporal scale to
598 establish the effect of those disturbance events.

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799 **TABLES**

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Site Code	Site Name	Location	Lat. (deg)	Long. (deg)	No. of visits	Haulout duration (hours) ($\bar{x} \pm SE$)	No. of individuals
BDH	Bagh an Da Dhoruis	Islay	55.93559	-6.15097	87	3.2 \pm 0.17	3
BHN	Bunnahabhainn	Islay	55.891175	-6.131105	123	5 \pm 0.31	7
BRP	Brein Phort	Jura	55.922896	-6.064843	23	5.3 \pm 2.64	3
CAS	Carragh an t-Struith	Jura	55.87061	-6.096444	4	2.3 \pm 0.93	2
CON	Colonsay North	Colonsay	56.1253	-6.1626	2	4.7 \pm 0.08	1
EGH	Eileanan Gainmhich	Islay	55.864512	-6.110327	59	3.9 \pm 0.36	6
EGR	Eilean Gleann Righ	Jura	55.968332	-5.986099	230	6.2 \pm 0.66	6
EST	Eileanan Stafa	South Uist	57.39659	-7.288119	35	6.9 \pm 0.63	1
HAU	Haun	South Uist	57.090523	-7.296631	8	3.5 \pm 0.76	1
HOU	Hough Skerries	Tiree	56.52	-7.020000047	1	0.6 \pm 0.00	1
HRT	Hairteamul	South Uist	57.084119	-7.229136	1	1.1 \pm 0.00	1
ISL	Nave Island	Islay	55.8991244	-6.34078397	1	0.5 \pm 0.00	1
RBL	Rubha Liath	Jura	55.962461	-5.950904	22	5.6 \pm 0.53	2
RBR	Rubha Bhoraraic	Islay	55.819718	-6.103997	4	1.6 \pm 0.87	3
SAN	Sanda Island	Kintyre	55.284856	-5.571027	4	2.9 \pm 0.86	1
SGB	Sgeiran a Bhudragain	Jura	55.958036	-5.946192	22	4.5 \pm 0.76	3

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803 **Table 1.** Listed are site code abbreviations for the site names of haulouts visited by telemetry tagged seals at locations around Islay.
804 Latitude, longitude coordinates define exact positions of haulouts. Also given are the number of visits, mean haulout duration and the
805 number of individuals that visited each site.

		Arrive			
		BHN	BDH	EGR	BRP
Depart	BHN	5			1
	BDH		1		
	EGR			6	
	BRP				
	BHN'	5		1	
	BDH'		1		
	EGR'			2	
	BRP'				

806

807 **Table 2.** Haulout/trip transition matrix showing where tagged seals departed from and
808 where they arrived and hauled out again after simulated disturbance trials. The total
809 number of disturbance trials resulting in each scenario are given. In the upper part of
810 the matrix (grey) are locations where seals hauled out again within the same low tide
811 period after being disturbed into the water. In the lower part of the matrix (pink)
812 are locations suffixed with ' , where seals hauled out again in any subsequent low tide period
813 having started a trip after being disturbed into the water. See Table 1 for full names of
814 abbreviated haulout locations.

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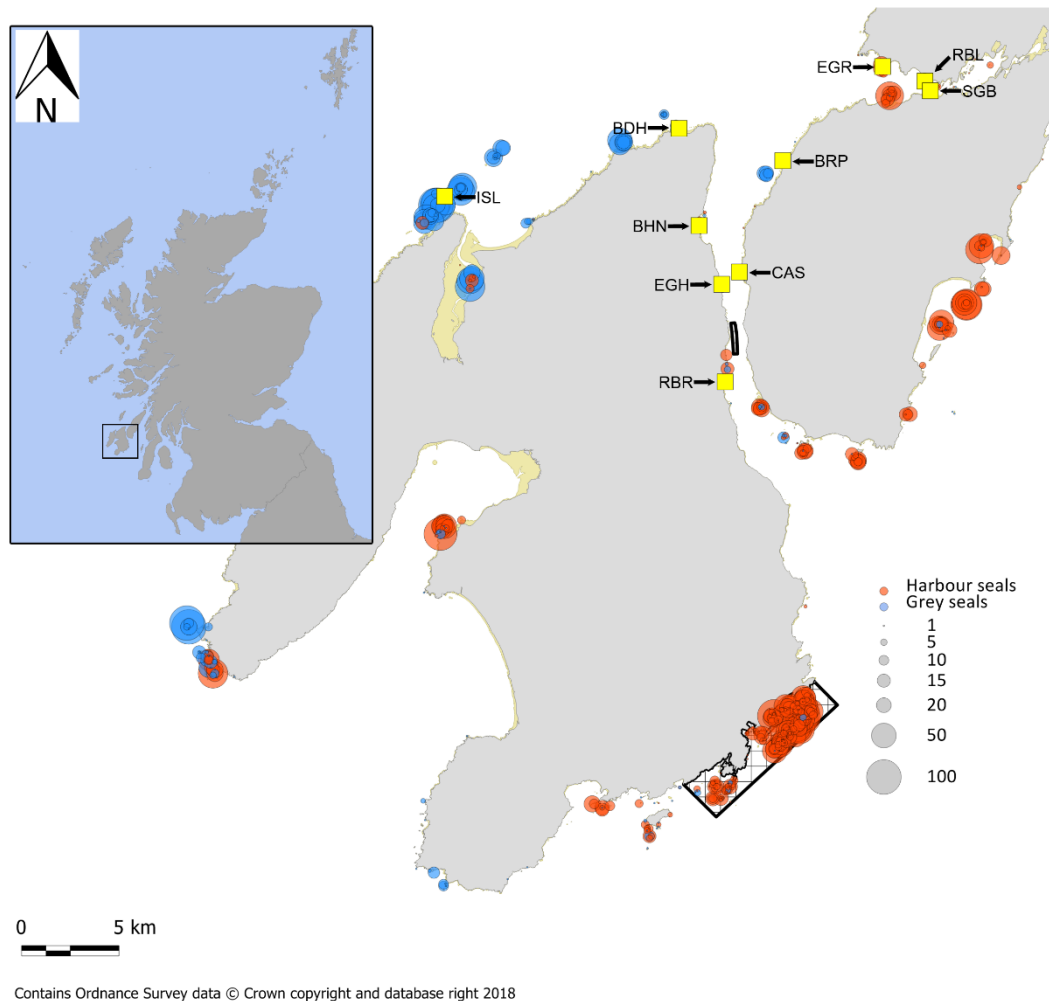
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822 **FIGURES**



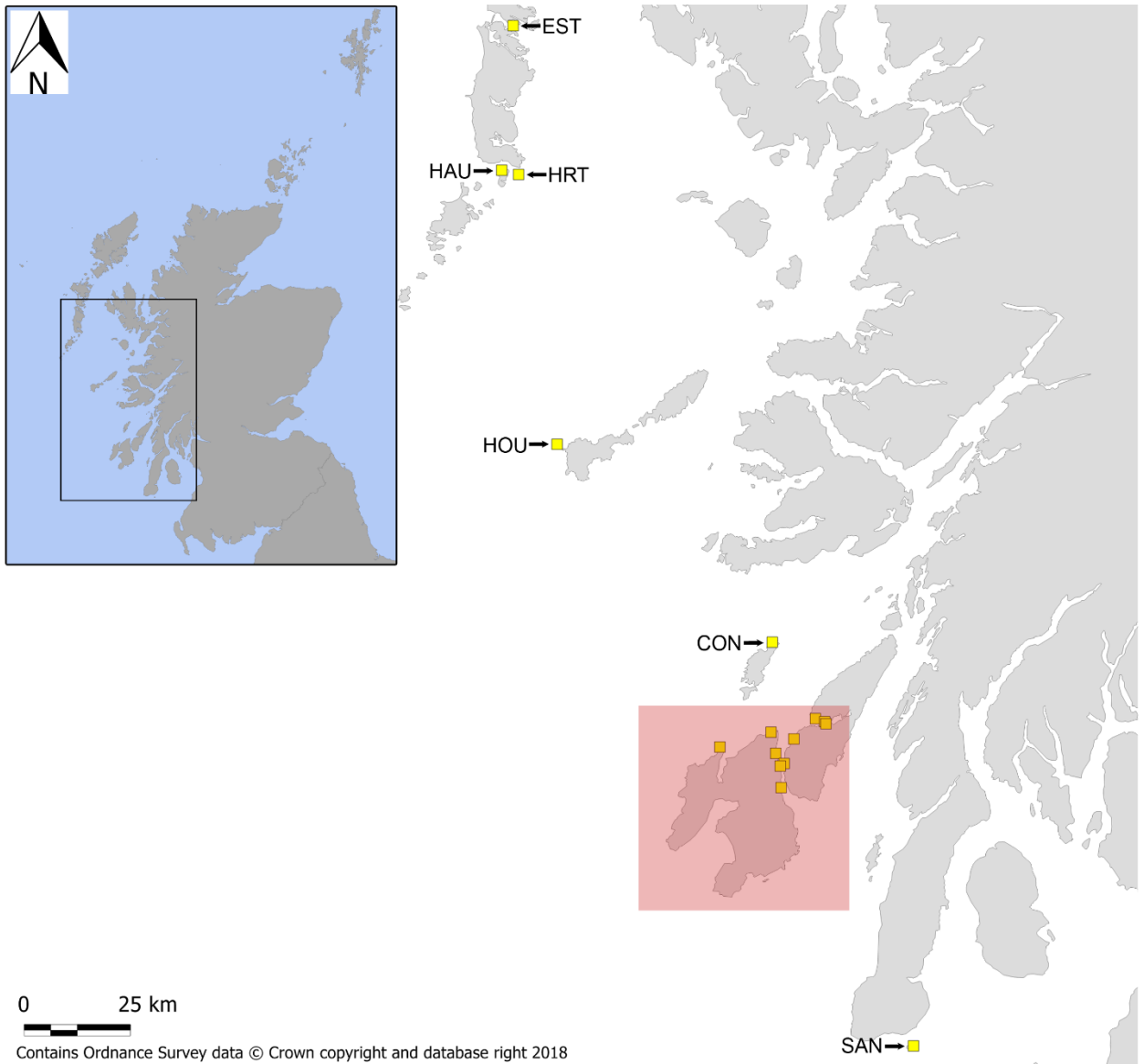
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824 **Figure 1.** The Sound of Islay and the South-East Islay Skerries SAC haulout sites. The
 825 South-East Islay Skerries SAC is delineated and shaded black. Boundaries of the
 826 proposed tidal turbine development within the Sound of Islay are also delineated in
 827 black. Yellow squares mark haulout sites visited by telemetry tagged seals in this study
 828 (See Table 1 for full names and latitude/longitude coordinates). Seal counts were taken
 829 from aerial survey data collected during the moult periods between 1990 and 2009. All
 830 aerial survey counts were carried out during a window of two hours either side of low
 831 tide.

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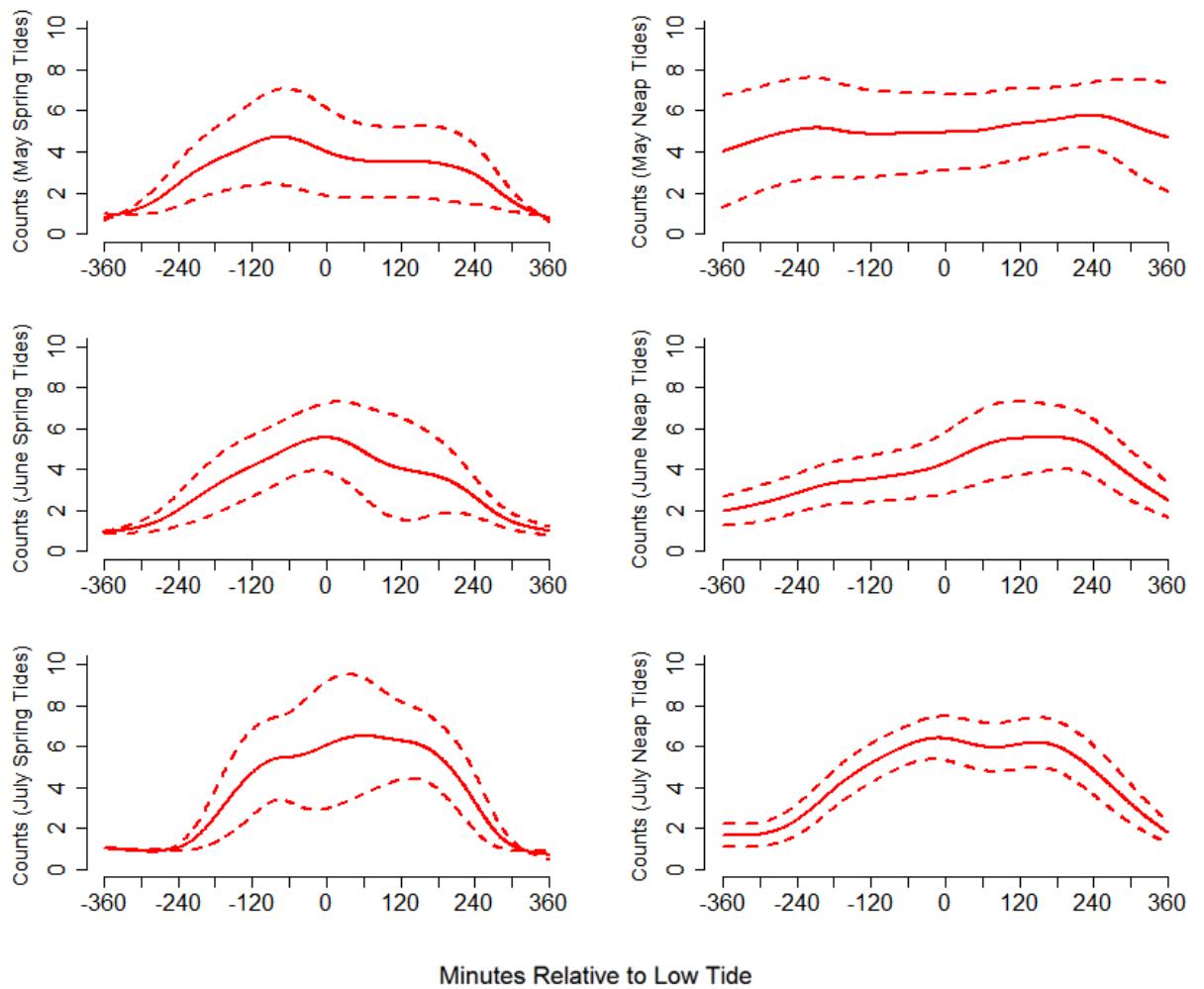
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 836 **Figure 2.** Wider geographical range of haulout sites visited by telemetry tagged seals
 837 marked by yellow squares (See Table 1 for full names and latitude/longitude
 838 coordinates). Haulout sites visited within close proximity of the Sound of Islay (pink
 839 shaded area) are presented in Figure 1.

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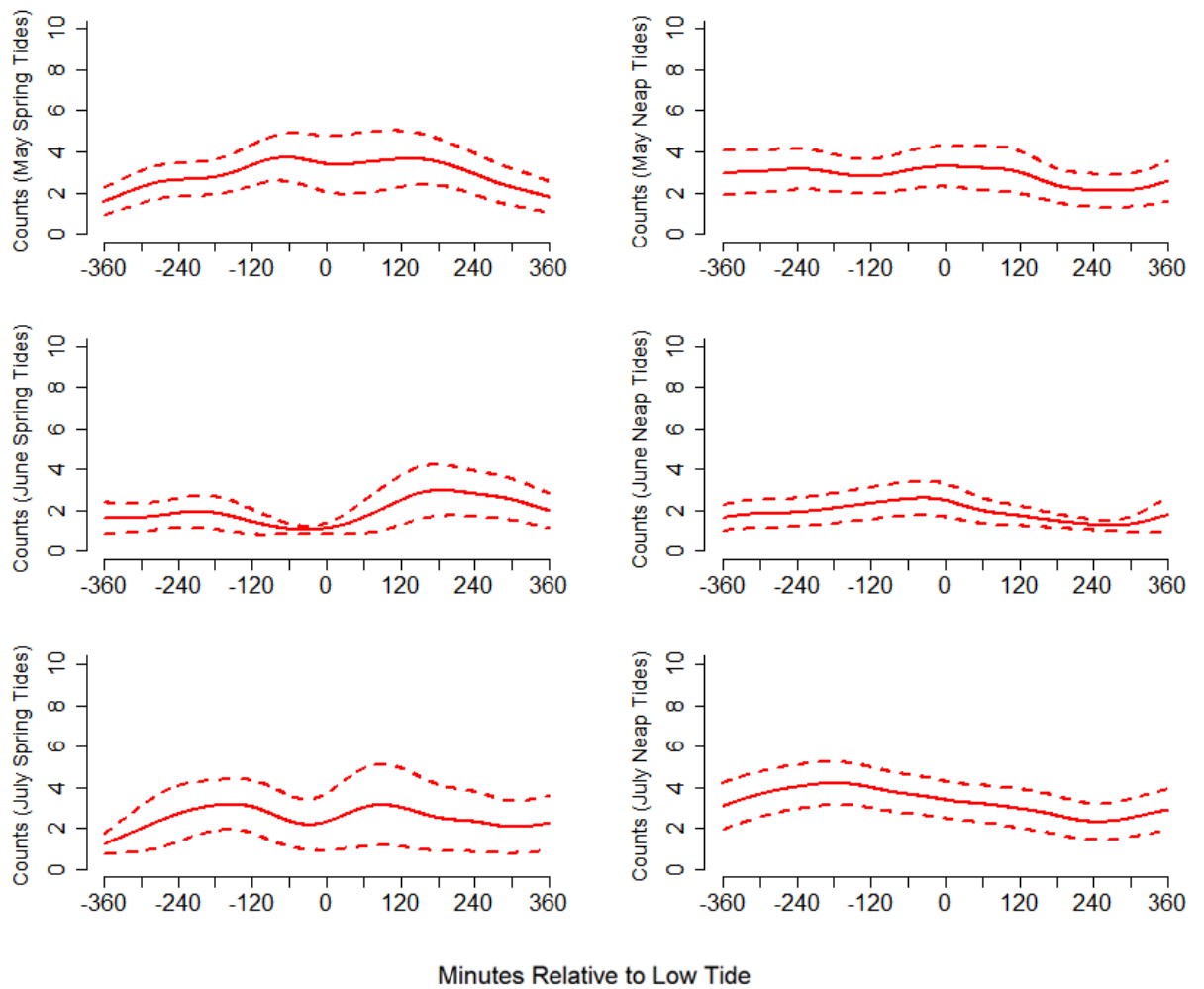


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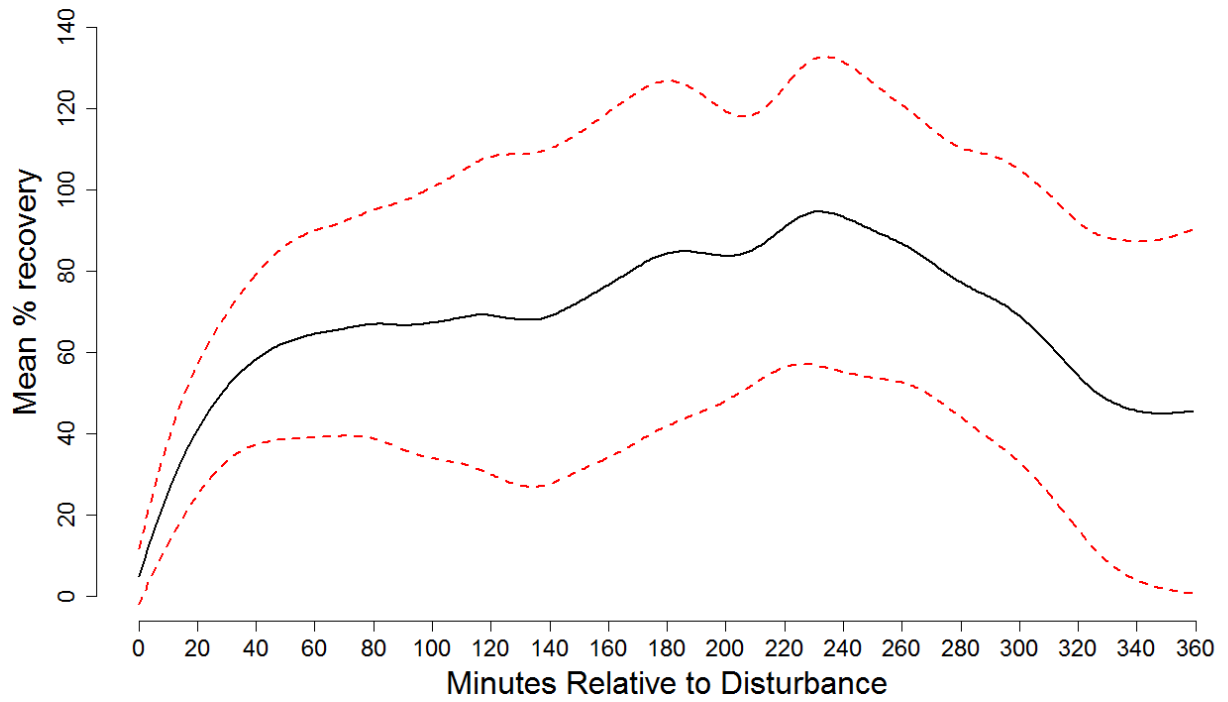
842 **Figure 3.** Mean counts of hauled out seals (solid red) with 95% confidence intervals
 843 (dashed red lines) with time relative to low tide at Bunnahabhain (BHN). Data are
 844 divided into spring and neap tide periods for May, June and July.

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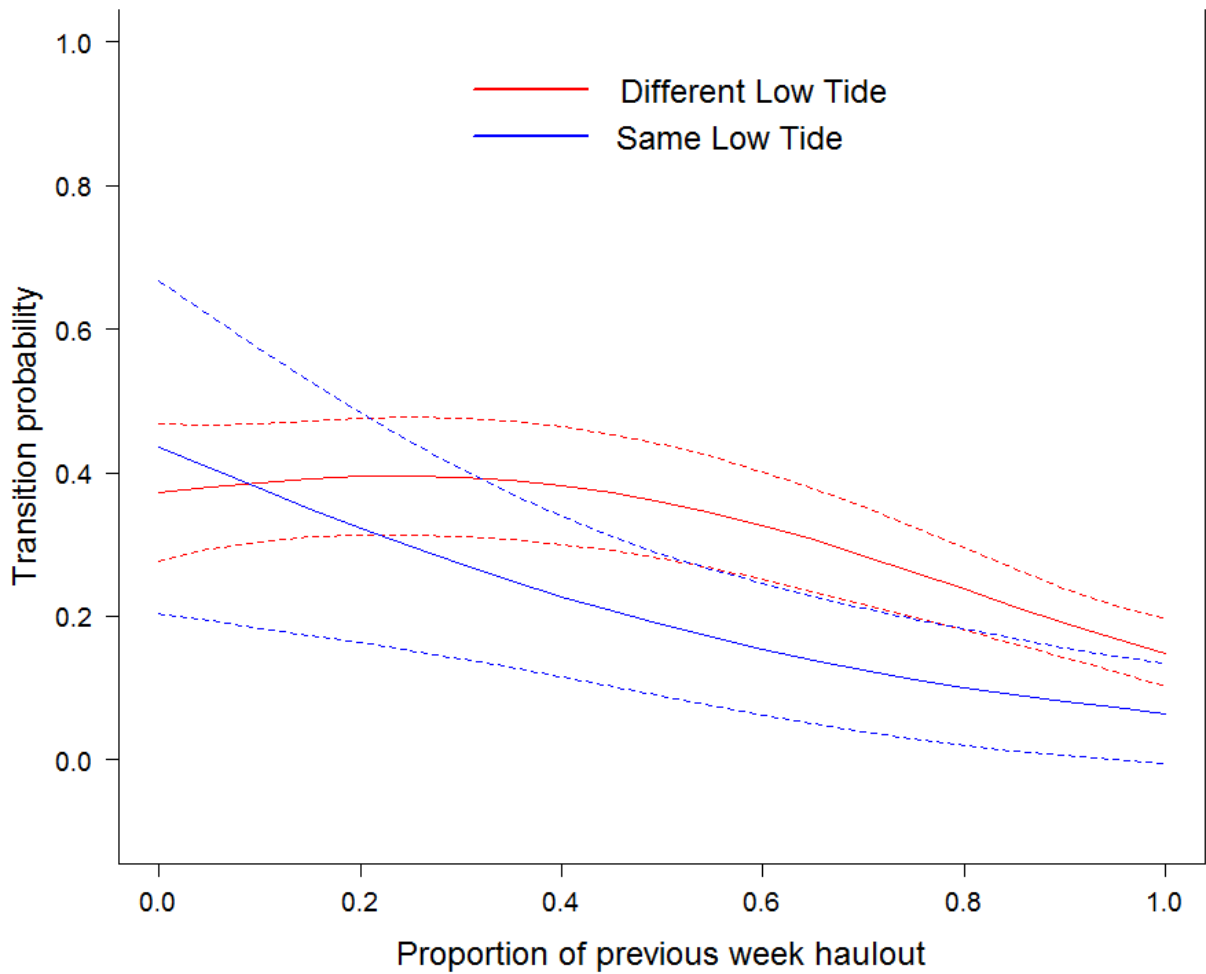


847
 848 **Figure 4.** Shown are the mean counts of hauled out seals (solid red) with 95%
 849 confidence intervals (dashed red lines) over minutes relative to low tide at Rubha
 850 Bhoraraic (RBR). Data are divided into spring and neap tide periods for May, June and
 851 July.



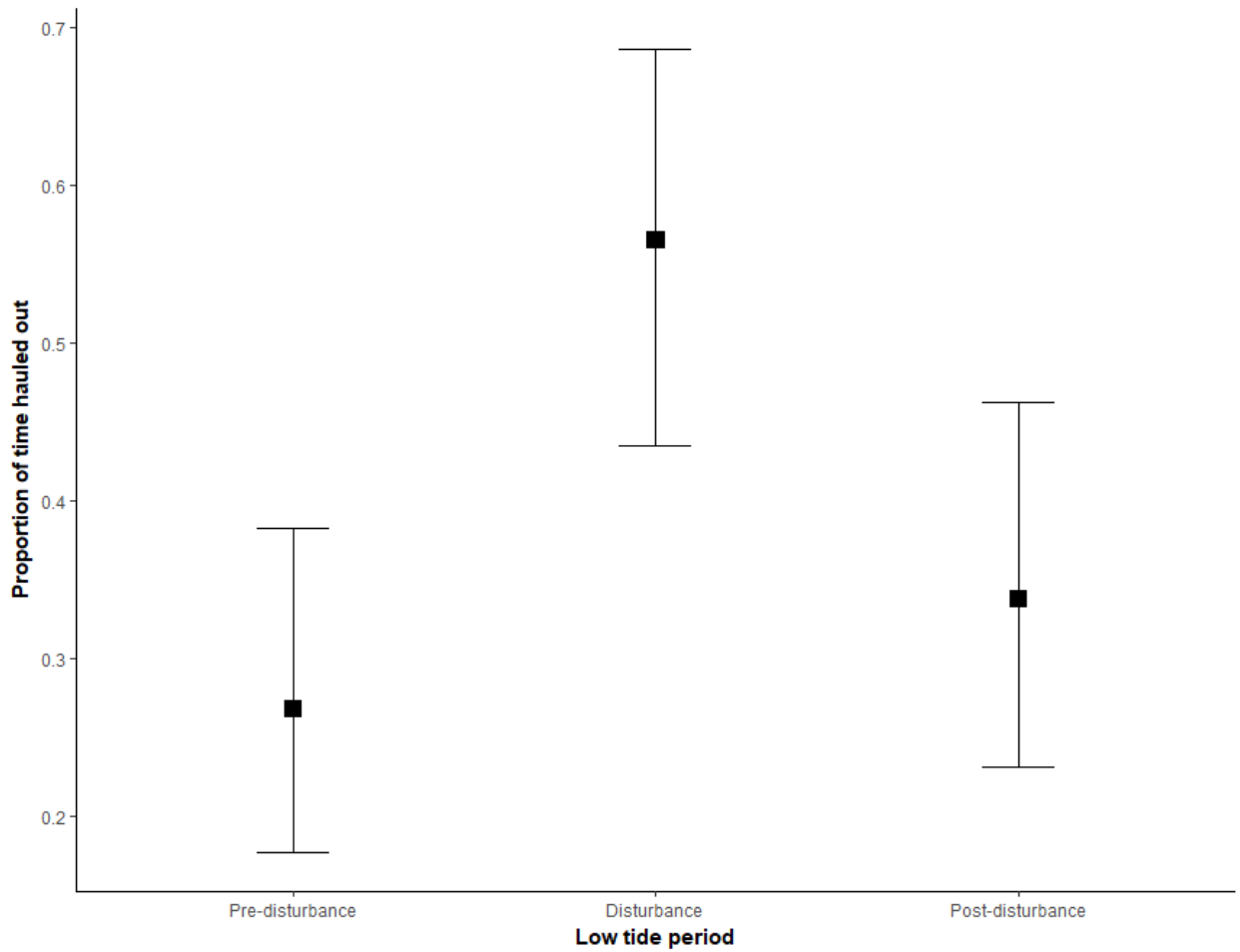
852

853 **Figure 5.** Mean percentage recovery of the number of hauled out seals (solid black
 854 line) with 95% confidence intervals (dashed red lines) against time (minutes) since
 855 disturbance trials. Data are for Bunnahabhain (BHN).



856

857 **Figure 6.** Transition probability i.e. having left a haulout site a seal then hauls out at a
 858 different haulout site (y-axis) is shown dependent on the proportion of haulouts in the
 859 previous week that were also at the haulout site a seal arrives at (x-axis). Transition
 860 probabilities are shown for the two scenarios of having ended a haulout a seal then
 861 hauls out again on the same (blue) or on a subsequent (red) low tide. Solid lines are
 862 model predictions with 95% confidence intervals as dashed lines.



863

864 **Figure 7.** GLMM model predictions of mean and 95% confidence intervals for the
865 proportion of time spent hauled out during pre-disturbance, disturbance and post-
866 disturbance low tide periods.

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